Section 9

Literature Review

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Literature review summary

Runoff from open lot livestock systems (beef and dairy) defined as CAFOs must be controlled by systems designed and managed to prevent the release of manure contaminated runoff for storms equal to or less than a 25-year, 24-hour design storm. This performance standard has been attained for open lot systems with some combination of clean water diversion, settling basins, runoff collection ponds, and irrigation systems (baseline system).

An alternative approach is to rely on overland flow and infiltration into cropland with perennial forage or grasses for treatment of open lot runoff. Such vegetative systems have been researched since the late 1960s. This paper reviews the research literature on VTSs for managing open lot runoff summarizing available science on system performance, design, and management.

Based upon this review of literature, the following conclusions are drawn about the application of VTS to manage runoff from open lot livestock production systems.

- Substantial research (approximately 40 identified field trials and plot studies) provides a basis for understanding the performance of VTS. A superior research knowledge base exists for performance of VTS as compared to baseline systems for CAFO regulation compliance.
- The baseline systems for CAFO regulation compliance perform well in the High Plains regions of the United States where significant moisture deficits exist (rainfall minus evaporation). However, the performance of these baseline technologies drops substantially for decreasing moisture deficits found in the central and eastern Corn Belt states. These trends have been established through computer modeling processes but not confirmed with in-field performance measurements.
- The existing research targeting VTS is confined to non-CAFO applications, likely due to past regulatory limits. Unique challenges exist in adapting these results and recommendations to CAFO applications.

• The pollutant reduction resulting from a VTS is based upon two primary mechanisms: sedimentation, typically occurring within the first few meters of a VTS and infiltration of runoff into the soil profile. Systems relying primarily on sedimentation only are unlikely to perform equal or better than baseline technologies. System design based upon sedimentation and infiltration is necessary to achieve a required performance level for CAFO application.

Introduction

The terms Vegetative Treatment System (VTS) and Vegetative Treatment Area (VTA) are used. VTA applies to a cropped area with perennial grass or forage specifically designed to manage runoff from an open lot livestock facility. VTS refers to the combination of treatment components including a VTA or Vegetative Infiltration Basin (VIB) and other possible treatment components (solids settling).

Runoff from open lot livestock production systems continues to be a contributor to surface water impairment. This literature review summarizes past research on VTSs when applied to open lot systems. This alternative technology may potentially achieve the same pollution control that is achieved by current EPA NPDES technology-based standard. A VTS has the potential for providing control of pollution from feedlot runoff that is functionally equivalent to the conventional impoundment and land application system for CAFO. The 2003 final Federal rule for the NPDES Permit Regulation and ELG and Standards for CAFOs (Federal Register 2003) states that for large CAFOs with dairy cows or beef cattle:

"(a) there must be no discharge of process wastewater pollutants into waters of the U.S. from the production area.

(1) Whenever precipitation causes an overflow of manure, litter, or process wastewater, pollutants in the overflow may be discharged into U.S. waters provided:

a) The production area is designed, constructed, operated and maintained to contain all manure, litter, and wastewater including runoff and the direct precipitation from a 25-year, 24-hour rainfall event;

b) The production area is operated in accordance with the additional measures and required by 412.37 (a) and (b) (note: defines management and record keeping expectations).

(2) Voluntary alternative performance standards. Many CAFO subject to this Subpart may request the Director to establish NPDES permit effluent limitations based upon site-specific alternative technologies that achieve a quantity of pollutants discharged from the production area equal to or less than the quantity of pollutants that would be discharged under the standards as provided by paragraph (a)(1)..." Part (1) sets the 25-year, 24-hour storm technology standard for baseline systems (runoff holding facilities dewatered by irrigation systems). Part (2) opens the door for alternative technology (such as a VTS) if they can be proven to achieve equal or less discharge of pollutants than the baseline technology (runoff holding pond plus irrigation). The site-specific comparison provision will place the burden of proof on the individual producer for comparing the baseline and alternative technology for individual farms.

Feedlot runoff characteristics

Most research defining the characteristics of runoff from open livestock systems was completed in the 1960s through the 1980s. Based upon this, research common characteristics have been published in accepted references from NRCS (table 9–1), Texas Agricultural Extension Service (table 9–2), and Experiment Stations of the North Central Regions land grant universities (table 9–3). Original data for many of these reported values is from Linderman and Mielke (1975); Gilbertson et al. (1979); Swanson et al. (1971); Gilbertson and Nienaber (1973); Gilbertson et al. (1975); and Gilbertson et al. (1972).

Runoff quality

Some generalizations about characteristics of feedlot runoff can be based upon this previously cited research:

• The solids fraction is roughly 10 times greater in runoff from snowmelt as compared to runoff from rainfall (table 9–3). Fields (1971) reported 2 to 2.5 times higher solids in snowmelt runoff as compared to rainfall runoff.

- Volatile solids (VS) typically represent about 50 percent or less of total solids in runoff.
- Approximately 40 to 80 percent of solids in runoff will settle in settling basins designed with 30 minutes or greater retention capacity.
- Increasing rainfall intensity leads to higher solids loss from the feedlot surface and greater VS or chemical oxygen demand (COD) concentration. Rainfall duration does not affect solids content of runoff.
- Ammonium and nitrate contents in the runoff decrease with continuing precipitation, indicating rapid leaching of these compounds from the feedlot surface.
- Phosphorus removal is closely related to solids removal and directly affected by rainfall intensity.
- Salt concentrations are the primary constituent of concern for crop performance that should be reviewed when runoff is used in land application.

| | | Runoff pond | | |
|-----------------|------------|-------------|--------|--|
| Component | Units | supernatant | Sludge | |
| Total solids | % w.b. | 0.30 | 17.20 | |
| Volatile solids | kg/1,000 L | 0.899 | 77.3 | |
| Fixed solids | kg/1,000 L | 2.10 | 94.4 | |
| COD | kg/1,000 L | 1.40 | 77.2 | |
| Nitrogen | kg/1,000 L | 0.20 | 6.19 | |
| Ammonium-N | kg/1,000 L | 0.18 | - | |
| Phosphorus | kg/1,000 L | _ | 2.10 | |
| Potassium | kg/1,000 L | 0.90 | 1.70 | |

Nitrogen content (kg N/1,000 L) of feedlot runoff at holding pond for:

| Annual rainfall | Below average conditions | Average conditions | Above average conditions |
|-----------------|--------------------------|--------------------|--------------------------|
| <64 cm | 1.6 | 0.49 | 0.26 |
| 64–89 cm | 0.26 | 0.13 | 0.066 |
| >89 cm | 0.066 | 0.044 | 0.022 |

Below average: No settling facilities between the feedlot and pond. Feedlot topography and other characteristics are conducive to high solids transport. High cattle density—more than 620 head/ha (250 head/a).

Average:Sediment traps, low-gradient channels, or natural conditions remove appreciable amounts of solids from runoff. Average
runoff and solids transport characteristics. Average cattle density—310 to 620 head/ha (125–250 head/a).

Above average: Highly effective solid removal, such as vegetated filter strips or settling basins that drain liquid waste through a pipe to storage pond. Low cattle density—less than 310 head/ha (125 head/a).

Table 9–2

Average runoff characteristics from beef cattle feed yards in the Great Plains (Sweeten 1991)

| Source | Total solids (ppm) | Electrical conductivity (mmho/cm) | Chemical oxygen demand (ppm) | Total nitrogen (ppm) | Total phosphorus (ppm) | Sodium (ppm) | Potassium (ppm) |
|-----------------------------|--------------------------|---|---------------------------------------|----------------------------|------------------------------|-----------------|--------------------|
| Feedlot runoff ¹ | | | | | | | |
| Average | 11,200 | 6,500 | 9,200 | 580 | 120 | 440 | 1,020 |
| Range | 3,000- | 3,200- | 2,200- | 80-1,080 | 50-300 | 230 - 590 | 340-1,320 |
| 0 | 17,500 | 8,600 | 17,800 | , | | | , |
| Pond effluent | , | , | , | | | | |
| South Texas | 2,500 | 4,500 | 1,100 | 180 | | 230 | 1,140 |
| Texas High Plains | , <u> </u> | 4,500 | 620 | 140 | 40 | 260 | 450 |

 $^1\,\mathrm{Seven}$ feed yards in TX, CO, NE, KS, and SD

| Fable 9–3 | Unpaved beef cattle feedlot runoff characte | eristics (Gilbertson et al. 1981) |
|-----------|---|-----------------------------------|
|-----------|---|-----------------------------------|

| Source | Total solids (%) | Volatile solids (%) | Chemical oxygen demand (ppm) | TKN (ppm) | Total phosphorus (ppm) | Electrical conductivity (mmho/cm) |
|----------|------------------------|---------------------------|---------------------------------------|--------------|------------------------------|---|
| Nebraska | | | | | | |
| Rainfall | 0.24-3.3 | 0.12 - 1.5 | 1,300-8,200 | 11-8,593 | 4-5,200 | _ |
| Snowmelt | 0.8 - 21.8 | 0.6 - 14.3 | 14,000-71,000 | 190-6,528 | 5-917 | 3–19 |
| Texas | 0.5 - 1.5 | 0.9 - 1.4 | 10,000-20,000 | 660-1,100 | 130-200 | 6-10 |
| Kansas | 0.8 - 1.9 | 0.36-0.96 | 800-16,000 | 165 - 1,580 | 9-242 | 2-13 |





Runoff quantity

Maps for estimating design storm and average monthly runoff volumes are available from chapter 10 of the Agricultural Waste Management Field Handbook (Soil Conservation Service 1992). Some common observations relative to volume of runoff from open lots include the following:

• A linear relationship exists between runoff volume and rainfall (fig. 9–1). A rainfall event greater than 1 centimeter is necessary for runoff to occur. An average prediction equation was suggested by Clarke et al. (1975):

Runoff (cm) = 0.56 x Precipitation (cm) – 0.84

- A greater slope for the prediction equation should be used in regions with lower moisture deficit (rainfall – evaporation). This would suggest that higher rainfall regions should expect greater runoff volumes for the same size storm, a factor that is not included in current predictive equations (Clarke et al. 1975).
- Feed yard slope and stocking rates have little influence on runoff amounts (Gilbertson et al. 1970 and Clark et al. 1975).
- Lots that are wet the previous day have less runoff than dry lots due to depressions created by animal activity creating more opportunity for water retention on wet lots (Clarke et al. 1975).

The volume of runoff from a feedlot for a given storm is commonly estimated using the NRCS Curve Number method. This method is described in the NRCS National Engineering Handbook, part 630 (Monkus 1964). For the purpose of estimating the volume of storm runoff from a feedlot, the following equation is solved for Q:

$$Q = \frac{\left(P - 0.2 \left(\frac{1000}{CN_{1}}\right) - 10\right)^{2}}{\left(P + 0.8 \left(\left(\frac{1000}{CN_{1}}\right) - 10\right)\right)}$$

where:

Q = volume of runoff (in)

P = rainfall (in)

CN₁= NRCS 1-d curve number

A CN_1 of 89 or 90 is commonly used for an unpaved feedlot, and a CN_1 of 97 or 98 is commonly used for a paved feedlot.

Pollutant mass in runoff

In addition to knowledge of volume and concentration, total mass of nutrient and solids in runoff can be useful in design of settling basins and land application sites. Nutrient mass balance data has been collected on a set of University of Nebraska research beef cattle feedlot pens over approximately a 5-year period (Erickson and Kissinger 2004) representing 120 separate pens of cattle over the entire finishing period. This data would suggest that runoff after settling will contain 27 kg total solids, 0.68 kg nitrogen and 0.32 kg phosphorus per finished animal (table 9–4). (Settling basins were designed to hold all runoff until after a storm event for the purpose of measurement of volume and collection of sample before release to a holding pond.)

| _ | Volume (L/finished animal) | Nitrogen | Phosphorus | Volatile solids | Total solids |
|----------------------------|-------------------------------|----------|----------------|--------------------|-----------------|
| | | | (kg/finished a | animai) | |
| Runoff | 3,600 | 0.68 | 0.32 | 13 | 27 |
| Standard deviation | 2,800 | 0.63 | 0.31 | 13 | 37 |
| Estimated total excretion | | 25 | 3.3 | 290 | 360 |
| % of excretion in runoff | 2 | 2.7 | 9.8 | 4.6 | 7.6 |
| Number of individual trial | $1s^1$ 120 | 112 | 48 | 80 | 64 |

Table 9-4Mass of solids and nutrients in runoff from beef cattle feedlot pens (Erickson and Kissinger 2004)

 1 One trial represents one pen of cattle entering the pen as calves or yearlings and fed to market weight. Feedlot is typically stocked at 30 m^{2} per animal with an average slope of 6%.

Performance of runoff collection ponds

Since runoff from open lots is weather dependent, most in-field monitoring efforts are challenged to collect data over a sufficient time period to accurately predict the long-term performance of control technologies. The only efforts to predict runoff holding pond performance identified in the literature were based upon performance models. No field studies were identified that provided field measurements of performance for runoff holding ponds based upon a 25-year, 24-hour storm event design criteria or other related criteria. It would appear that once the EPA established their technology based ELG, no efforts have been made to document in-field performance of these design criteria.

Planning software titled Animal Waste Management (AWM) is maintained by NRCS and commonly used for sizing of manure storage and runoff holding ponds (Wilson et al. 2003). An evaluation of the storage sized by AWM was compared against a water-balance model for storages using 30 years of weather data for 10 United States sites (Moffitt et al. 2003). The comparison revealed that 0 to 73 percent of the 30 years produced events requiring land application at shorter intervals than the design critical storage to maintain an acceptable storage volume for a 25-year, 24-hour storm. If pump down during these periods did not occur, spillway flow would result during 0 to 40 percent of the modeled years. Management decisions during these periods when storage capacity was inadequate and sizing of the de-watering pump were two critical factors minimizing spillway flow.

A computer model developed by Kansas State University (Koelliker et al. 1975) predicts the portion of runoff controlled by a conventional runoff holding pond and irrigation system (sized to pump 10 percent of the holding pond volume per day). This model was used to evaluate a basin system for five Kansas sites and predicted that such systems perform better in more arid climates (table 9–5). Full (100%) control was predicted in southwest Kansas while only 93 percent control (and 47 days of discharge over 30 years) was predicted for northeast Kansas. Discharges most commonly resulted from a series of precipitation events less than the design storm over an extended period of time when land application of liquid was judged to be not feasible (saturated soil conditions in land-application site).

An Iowa State University application of the Kansas State model (Wulf et al. 2003, 2004) provides additional support for the Kansas State observations. Based upon Iowa Department of Natural Resources minimum design criteria, five alternative design and management scenarios were modeled with 50 years of weather data for six Iowa locations. The resulting predictions suggested that between 70 and 90 percent of runoff could be controlled based upon a 25-year, 24-hour storm design criteria with additional normal runoff storage requirement mandated by Iowa regulations. (States may require storage capacity in addition to the minimum Federal ELG requirement of a 25-year, 24-hour storm capacity. This additional capacity is typically sized to address average runoff over a pre-determined time. Iowa has established five methods for estimating this capacity based upon the planned schedule for dewatering of the holding pond.) The every event pump out results (table 9-6, col. 2 and 3) compare favorably with the Kansas State results.

Table 9–5Performance of runoff control facility sized to hold runoff from an unsurfaced feedlot for a 25-yr, 24-h precipita-
tion event as evaluated over a 30-yr period (Koelliker et al. 1975)

| Location | Runoff control (%) | Years with overflow | Avg. number of days with overflow ¹ | Number of days with discharge over 30 years |
|--------------|-----------------------|---------------------|---|---|
| Northwest KS | 98.6 | 2 | 1.5 | 3 |
| Southwest KS | 100.0 | 0 | 0 | 0 |
| Central KS | 97.9 | 3 | 2.3 | 7 |
| Southeast KS | 95.5 | 9 | 3.6 | 32 |
| Northeast KS | 93.0 | 9 | 5.2 | 47 |

¹ During years with overflow

The predicted performance of the baseline system illustrated regular discharge occurrences for all scenarios evaluated. Northeast and East Central Iowa conditions produced the most frequent discharges and the lowest volume of runoff control. Land application systems that were not able to land apply runoff following each precipitation event were more likely to have discharge. Increasing volume of storage provided some reduction in runoff control but did not eliminate discharges (fig. 9–2). The baseline system currently defined in the ELG (Federal Register 2003) performs well under High Plains regional conditions, as found in western Kansas, but not nearly as well in regions with higher precipitation levels, extended wet periods, or less conducive to use of pivot irrigation systems. To improve runoff control, it was further identified that extending the season for land application in the spring and fall produced the greatest benefits (extended pump out period results in table 9–6). Increasing pumping rate by 2.5 times or increasing storage capacity by 10 percent produced only minor improvements in increased runoff control (Wulf et al. 2003). Figure 9–2 illustrates the value of additional storage for a Central Iowa feedlot. Increasing total pond capacity from 30 to 48 centimeters (12–19 in) of total runoff produced a reduction in the runoff control, but did not eliminate discharges.

Table 9–6

Performance of runoff control facility sized to hold runoff from an unsurfaced feedlot designed based upon Iowa Department of Natural Resource criteria and evaluated over a 50-yr period (Wulf et al. 2004)

| | Every event pump out | | April and l | April and Nov. pump out | | Extended pump out period | |
|---------------------------|----------------------|----------|-------------------|-------------------------|-------------------|--------------------------|--|
| | Runoff control | Overflow | Runoff control | Overflow | Runoff control | Overflow | |
| Location | (%) | (d/yr) | (%) | (d/yr) | (%) | (d/yr) | |
| Northwest IA | 90.1 | 2.7 | 78.0 | 7.7 | 88.5 | 3.7 | |
| Southwest IA | 88.5 | 4.1 | 72.4 | 10.4 | 83.7 | 6.7 | |
| Central IA | 87.6 | 3.8 | 77.7 | 9.2 | 87.2 | 5.3 | |
| Southeast IA | 90.1 | 3.9 | 79.2 | 8.8 | 83.7 | 6.7 | |
| East Central IA | 82.3 | 6.1 | 64.5 | 13.4 | 80.3 | 7.8 | |
| Northeast IA | 81.3 | 6.0 | 66.5 | 12.9 | 87.3 | 5.6 | |
| Basin capacity – | | | | | | | |
| Amount of runoff 10–12 cm | | 20-25 | 20–25 cm | | 20–25 cm | | |





A second Kansas State University study used the Koelliker model to estimate the baseline system volume necessary to provide 100 percent control of runoff based upon weather records for a 25-year period (Anschutz et al. 1979). The volume of the holding basin varies substantially with location, as illustrated in table 9-7. A holding pond for the same size feedlot will be between 3 and 6 times larger in the central and eastern Corn Belt as compared to western Kansas. This assumes that the all locations would have access to dewatering capacity equal to a pivot application system. Such systems are less commonly found in many regions outside of the High Plains states. With other land application methods, additional storage capacity would be needed to compensate for the slower dewatering rates. The study further observed a low correlation ($r^2=0.33$) between a 25-year, 24-hour storm design criteria for pond sizing and the estimated "nodischarge" pond size based upon 25-year weather records. Moisture deficit was better correlated ($r^2=0.80$) to the "no-discharge" pond size.

VTS performance

Performance models for VTS

An Iowa State University VTS software modeling tool is designed to predict the performance of a site-specific VTS to meet the Voluntary Alternative Performance Standards of the new EPA CAFO rules (Wulf et al. 2004). The VTS model performs site-specific modeling using daily weather inputs to estimate the performance of site-specific feedlots and VTS designs. The model is run for each of 25 weather years so that the performance of the alternative VTS (median outflow for 25-year period times pollutant concentration) can be compared to the performance of a baseline containment system at the same site following the procedures outlined by the Voluntary Alternative Performance Standards provisions of the CAFO regulations (Federal Register 2003). At the time this literature review was published, the model verification process was complete and the model was approved by the EPA.

Table 9-7

Relative size of runoff holding pond and land application system capable of pumping 2,850 L/min or 750 gpm during all seasons. Holding pond is sized to avoid all discharge based upon 25 years of weather data (Anschutz et al. 1979).

| Location | Pond volume m ³ (10 ⁶ gal) | Relative size to Garden City, KS | Location | Pond volume m ³ (10 ⁶ gal) | Relative size to Garden City, KS |
|------------------|--|--|---------------------|--|--|
| Garden City, KS | 17,376 (4.6) | 1.0 | Wooster, OH | 226,853 (60.0) | 13.0 |
| Sacramento, CA | 57,760 (15.3) | 3.3 | Minneapolis, MN | 56,374 (14.9) | 3.2 |
| Dublin, GA | 110,936 (29.3) | 6.4 | Oklahoma City, OK | 38,771 (10.2) | 2.2 |
| Boise, ID | 19,980 (5.3) | 1.1 | Centerville, SD | 51,478 (13.6) | 3.0 |
| W. Lafayette, IN | 103,946 (27.5) | 6.0 | Hereford, TX | 23,998 (6.3) | 1.4 |
| Urbana, IL | 62,968 (16.6) | 3.6 | College Station, TX | 54,761 (14.5) | 3.1 |
| Independence, KS | 37,186 (9.9) | 2.1 | | | |

Several Minnesota agencies have collaborated to develop a systematic procedure to identify appropriate applications of VTSs to feedlot runoff (Brach 2003; Minnesota Pollution Control Agency 2003). They have developed a standard identifying five levels of control (including VTA) and appropriate application of those five levels to individual situations based upon farm size and proximity to water. The team has developed a model, FLEVAL: An Evaluation System to Rate Feedlot Pollution Potential, to objectively evaluate feedlot pollution potential (*http://www.bwsr.state.mn.us/outreach/engineering/fleval.html*). Overcash et al. (1981) describes an additional model for predicting performance of a vegetative system located down-gradient from a manured land application site.

Solids removal performance

Solids removal via settling basins has been investigated for swine and bovine open lot runoff. Early studies of settling by Moore et al. (1973) using Imhoff cones showed that the majority of solids from beef feedlots settled within 10 minutes. From 10 minutes to 100 minutes only a slight improvement in settling was found. Fischer et al. (1975) concluded that the settling characteristics of hog manure are highly variable, but most settling occurs within the first 100 minutes. More recently Lott et al. (1994) examined solids in manure from Australian feedlots and differentiated two components: large particles that settled within 10 minutes and small particles that required extremely long settling times. The rapidly settling portion varied from 45 to 75 percent of the total solids. Sedimentation basin design based upon a maximum settling velocity of 0.003 m/s was recommended by Lott et al. (1994).

A 2-year study of settling basin performance below a swine facility and a beef feedlot in Iowa was conducted in the early 1990s (Lorimor et al. 1995). Solids in the swine runoff were reduced 29 percent from 3.1 percent to 2.2 percent wet basis. Solids concentration in the retained solids within the basin increased to an average of 12.7 percent. On a mass basis, the settling basin below the swine lot retained an average of 46 percent of the solids, 31 percent of the total Kjeldahl nitrogen (TKN), and 31 percent of total phosphorus (P) over the 2 years of monitoring. Settling below the earthen beef feedlot in this study removed a mean of 64 percent of the total solids, 84 percent of the TKN, 80 percent of the total P, and 34 percent of potassium (K).

Woodbury et al. (2003a) reported total nitrogen mass reduction of about 45 percent for a settling basin on a central Nebraska beef cattle feedlot over a 2-year period. Gilbertson and Nienaber (1973) observed that 71 percent of total solids that eventually settle will do so in the first 15 minutes representing 40 percent of total solids in runoff (Gilbertson et al. 1972).

Gilbertson et al. (1971) reported on performance of a batch system and a continuous-flow system for feedlot runoff. The batch system was more efficient in solids removal but suffered from management challenges including removal of settled solids. Dual settling basins were recommended to encourage greater drying and simplified solids management with solidshandling equipment. A continuous-flow system consisting of three porous dams in a settling channel recovered 50 percent of the total solids with 80 percent settling behind the first damn. Cold-weather solids settling proved a greater challenge, with solids remaining in a suspended form for longer periods at near-freezing temperatures. Only 42 percent of total solids were captured by the continuous-flow system during winter thaws.

Over a 2–1/2 year period, Swanson and Mielke (1973) monitored a broad, flat channel with two or three galvanized hardwire meshes installed to settle solids from runoff. It was estimated that 80 percent of the total solids were removed during the period observed. Key design recommendations included:

- channel length at least 6 times the channel width
- channel depth should exceed screen height to permit emergency overflow
- first screen placement at to half the length of channel from the inlet with additional screens equally spaced
- solids depth maximum of 38 centimeters (15 in)
- inclusion of a hard-surface channel bottom to facilitate equipment operation

The first component of any open feedlot runoff treatment system, whether it is total-containment system or alternative technology, should be solids settling, as is currently required by many state laws. Properly designed and managed solids settling basins should remove about 30 percent of the N and P from the runoff from swine lots and up to 80 percent of each from bovine lot runoff. Design recommendations for solids settling basins are available from MWPS (1985); Gilbertson and Nienaber (1973); and Sweeten (1991).

VTA performance

The author uses the terms VTA or vegetative treatment areas to represent the same technologies often referred to by other authors as vegetative filter strips. The author's choice of terminology differentiates VTAs applied to open lot livestock facilities from vegetative filter strips commonly used down gradient of cropland. Although both technologies share some similarities, there are distinctive differences in design and management.

This review of the literature assembled performance data from 16 research citations reporting 40 sets of performance data under field conditions (table 9–8) and an addition 17 research citations reporting 61 sets of performance data under simulated conditions (table 9–9). These research results are for both VTAs and VIBs. The preponderance of the performance data is for a VTA. VTA efficiency is estimated in the literature by comparing the reduction of pollutant concentration and/or mass entering and leaving the VTA. Pollutants of concern in livestock runoff include solids, nitrogen, phosphorus and pathogens. In addition, summaries of performance observations beyond specific pollutant reductions are reported in table 9–10.

Ikenberry and Mankin (2000) defined a VTA as a band of planted or indigenous vegetation situated downslope of cropland or animal production facilities that provides localized erosion protection and contaminant reduction. Planted or indigenous vegetation is defined as pasture, grassed waterways, or cropland that is used to treat runoff through settling, filtration, adsorption, and infiltration. Murphy and Harner (2001) identified four primary approaches used in VTAs:

- VTAs should be designed with a 1 to 4 percent slope and 61 meters (200 ft) of filtering length per 1 percent slope. Total area should be designed to match crop nitrogen uptake with estimated N in runoff. Uniform flow across filtering slope is necessary, typically requiring laser-guided land leveling equipment.
- Constructed wetlands have been applied to open lot runoff. Design and management is challenged by the intermittent flow from open lots. The authors suggests that seasonal open lots used for winter livestock housing and empty during the summer may be a preferred system for constructed wetlands.
- Infiltration basins are a containment type of system with a 30 to 60 centimeters (12–24 in) berm place around the vegetated area. They can be de-

signed as discharge or non-discharge systems. Infiltration area necessary to infiltrate design runoff within 30 to 72 hours must be considered in sizing of infiltration basin area.

• Terraces, similar to infiltration basins, have been used to contain runoff on sloped areas. Both overflow and cascading terraces have been used. Overflow terraces move runoff from one terrace to an adjacent terrace at a lower elevation by cascading of runoff over the terrace top or by plastic tile drains. Serpentine terraces move runoff back and forth across the face of a slope. In both situations, the upper terrace is typically used for solids settling.

VTAs provide an opportunity for reduction of pollutants in runoff through two primary mechanisms: sedimentation, typically occurring within the first few meters of a VTA, and infiltration of runoff into the soil profile (Pope and Stolenberg 1991). The soil system also provides a physical structure and biological environment for treatment of pollutants including filtration (restricting movement of most protozoa and bacteria), immobilization (soil cations immobilizing ammonium), aerobic processes (conversion of organic compounds to water and carbon dioxide), and anaerobic process (conversion of nitrates to nitrogen gas). The VTA also allows the recycling of nutrients by plants (Fajardo et al. 2001).

VTA flow can be classified as either channelized or uniform flow (Dickey and Vanderholm 1981a). Their work showed that "the channelized flow system required a flow length over five times longer than the overland flow systems to achieve a similar concentration reduction." Dillaha et al. (1988) studied concentrated flow effects on removal efficiencies and found that lower removal efficiencies occurred in VTAs with concentrated flows than in VTAs with shallow, uniform flow.

Surface flow in channelized-flow VTAs concentrates into channels. One can more clearly define these as gullied or preferential-flow systems. If gullied or preferential flow develops, non-uniform loading of VTA will reduce performance of the system due to soil erosion and reduced utilization of the VTA area. Uniformflow systems allow a uniform loading of waste (across the width of the VTA) at a relatively shallow depth (<4 cm). Uniform depth across the entire width of the VTA results in a slower velocity through the system, allowing sediment and nutrients to be trapped by the vegetation and adsorbed by the soil, and ultimately more efficient removal of nutrients and sediment from the waste stream. Dickey and Vanderholm (1981b) showed progressively better removal of N and ammonium (NH₄⁺) over 100 meters (300 ft) of overland flow in a VTA for a 100 head dairy and 500 head beef lot as shown in figure 9–3. Lim et al. (1997) and Chaubey et al. (1995) demonstrated a first-order exponential relationship better described the interaction between VTA length and pollutant transport. Data from 10 separate studies conducted over the last 25 years (fig. 9–4) show that 80 percent reductions of TKN and total P are achievable as a function of the ratio of VTA area to the feedlot drainage area.

Solids removal—Extensive research has been conducted on solids removal by VTA. Total solids are commonly reduced by 70 to 90 percent (tables 9–8 and 9–9). Variations occur due to site-specific conditions such as vegetation; slope; soil type; size and geometry of filter strip; and influent solids concentration. When receiving runoff directly from a feedlot, VTAs remove most solids within the first few meters of the filter strip. Coyne et al. (1998) found most reductions in con-





centration occurred in the first 4.5 meters. Chaubey et al. (1995) showed improved P removal effectiveness from swine lagoon effluent with increased VTA length up to 9 meters (30 ft). Solids reduction would likely perform in a similar manner. Chaubey et al. (1995) noted that removal of total suspended solids and chemical oxygen demand in VTA increased for lengths up to 3.1 meters. This quick reduction can be attributed to a significant reduction in flow velocity due to vegetation retarding the flow and soil conditions conducive to infiltration.

Fecal coliform removal—More research on fecal coliform (FC) removal by VTAs is needed. Reported values vary greatly and few studies have been conducted on large scale VTAs. Fajardo et al. (2001) report FC removal rates between 64 percent and 87 percent when using small-scale simulated runoff events with stockpiled manure. Lim et al. (1997) found that all FC were removed in the first 6.1 meters of a VTA used to treat runoff from a simulated pasture. Average FC removal in the studies reported was 76.6 percent (Ikenberry and Mankin 2000). A model for describing fecal pathogens in vegetative filter strips was being assembled by Zhang et al. (2001) and linked to an existing model of VTA hydrology and sediment transport, although data were not available to test the model at the time this research paper was prepared.

Figure 9–4 Nutrient removal by VTA based upon VTA to feedlot drainage area (DA) ratio for references listed in tables 9–8 and 9–9



Table 9-8 Summary of VTA performance when placed on commercial or research livestock facilities

| Study descrip | tion | | | | VTA information | | | | | | |
|---|--|-----------------|--|-------------------|-----------------|------|--------------|---|--|--|--|
| Reference | Summary | Study period | Pollutant source | Settling basin | Length (m) | | Slope (%) | Vegetation | Soil | | |
| Baker and Young 1984 | Milking center wastewa- ter and open lot runoff from a 54 cow dairy was directed to settling basin and VTA. Four earthen berms located at 9 m in- tervals were designed to | 5/82 – 5/84 | Milking cen- ter waste- water only | Yes | 91 x 23 | | 10 | Orchardgrass and foxtail at upper end. Hairy crabgrass in drier areas | VTA only VTA+basin VTA only VTA+basin | | |
| | create a cascading type system. System was mon- itored over 2 yr | | Milking cen- ter waste- water and paved dairy lot runoff | Yes | 91 x 23 | | 10 | Orchardgrass and foxtail at upper end. Hairy crabgrass in drier areas | VTA only VTA+basin VTA only VTA+basin | | |
| Dickey and Vanderholm 1981a | Four different VTA sys- tems after settling basins at actual feedlots | 17 mo | Dairy farm | Yes | 91 | 1.00 | 0.5 | Reed canary, bromegrass, and orchard- grass | | | |
| *Influent conc | entrations estimated from a | similar site | 450 head beef feedlot | Yes | 61 | 0.70 | 2 | Fescue-alfalfa mix | sandy | | |
| *Channelized | flow VTA (serpentine terrac | e channel) | 500 head beef feedlot | Yes | 533 | | 0.25 | | | | |
| *Vegetated ter | race channel and grassed w | aterway | 480 head swine fin- ishing fa- cility | Yes | 148 | | 0.25 | Garrison creep- ing foxtail | | | |
| Fausey et al. 1988 | Infiltration basin used with 56 head of beef cat- tle on concrete lot | 3 yr | 56 head beef feedlot | Yes | 6 x 27.5 | 0.7 | 1 | Reed ca- narygrass 1) Drain tile with slope 2) Drain tile across slope | silt loam | | |
| Edwards et al. 1986 | Infiltration basin used with 56 head of beef cat- tle on concrete lot | 3 yr | 56 head beef feedlot | Yes | 6 x 27.5 | 0.7 | 1 | Reed ca- narygrass, 1) VTA and settling basin 2) VTA only | silt loam | | |
| Harner and Kalita 1999; Keaton 1998 | 300 head feedlot runoff is directed to set- tling basin and VTA, | 2 yr | 300 head beef feed- lot | Yes | 427 | 0.97 | 0.3–4 | Bromegrass | silty clay loam | | |
| | 300 head beef feedlot discharges to VTA | 2 yr | 300 head beef feed- | Yes | 239 | 0.23 | 0.5–2 | Bromegrass | sandy loam | | |
| | Both facilities are in Kansas | | lot | | | | | | | | |

Table 9-8 Summary of VTA performance when placed on commercial or research livestock facilities—Continued

| TS | TSS | BOD5 | COD | Total N | TKN | NH ₄ -N | NH3-N | NO ₃ -N | Total P | Ortho- P | FC 2/ | FS 2/ | E. Coli | ** |
|-----------------------|----------|------|------------------------|-------------------------------|------------------------|----------------------|---|--------------------------------|------------------------|----------------------|----------|----------|------------|---|
| 90 95 99 100 | | | 96 98 100 100 | 97 99 100 100 | 97 99 100 100 | | 99 98 100 100 | 82 81 99 99 | 98 98 100 100 | | | | | c ^{3/} c m ^{3/} m |
| 45 65 97 | | | 56 65 98 | 46 60 97 | 46 60 97 | | 55 40 98 | $-68 \\ -17 \\ 92$ | 68 68 98 | | | | | c c m |
| 98 73.1 | | | 98 85.4 | 98 | 98 80.1 | | 97 86.2 | 94 | 98 78.2 | | | | | m c |
| 63.1 | | | 81.2 | | 71.1 | | 71.5 | | | | | | | с |
| 79.7 | | | 92.1 | | 83.1 | | 83.4 | | | | | | | c c |
| 61–81 55–83 | | | 69–87 59–86 | Org N 69–85 59–87 | | 69–92 56–89 | NO ₃ -N b 1 pp After 1 a 76 and 6 | efore: m and 2: 4 ppm | 62–91 63–89 | 73–93 67–90 | | | | c c |
| 82 80 66 61 | | | 85 83 69 65 | Org N 80 78 70 66 | | 50 50 73 72 | | -643 -940 -733 -1150 | | 80 74 77 70 | | | | c m c m |
| | 65 76 | | | 26 50 | | 44 63 | | 2 34 | 14 42 | 18 45 | | | | c m |
| | 78 83 | | | 73 59 | | 74 74 | | 95 87 | 71 52 | 64 44 | | | | c m |

Percent reduction

Table 9–8

Summary of VTA performance when placed on commercial or research livestock facilities-Continued

| | Study descript | ption | | | VTA info | ormatio | 1 | | | |
|--|--|---|---|------------|---------------|---------|--------------|---|------------------|---|
| Reference | Summary | Study period | Pollutant source | | Length (m) | AR | Slope (%) | Vegetation | I | Soil |
| Komor and Hansen 2003 | Settling basin and VTA were placed below two cattle feedlots and mon- itored for seven storm events | 1995–96 | 200 head ca- pacity lot (35 cattle during test) 225 head foodlot | Yes Yes | 79 58 | 0.2 | 1.2 0.5 | Grass | | silt loam loam |
| Lorimor et al. 2003 | Runoff from con- crete open lot beef facility is directed to settling basin, to- tally bermed infil- tration basin (IB), and constructed wetland (CW) | 1997 to present —data based upon 5 yr | 380 head concrete beef cattle facility | Yes | 108 | 0.18 | 0 | IB - Reed canarygrass CW–Com-m tails | s non cat- | Loam IB: IB + CW: IB + CW: |
| Mankin and Okoren 2003 | 300 head heifer feedlot with runoff directed to settling | May 2001–May 2002 | 300 head dairy heifer feedlot | Yes | 150 | | 2 | Fescue | | silt loam |
| | basin (stage 1) and VTA (stage 2) | | leculot | | | | | Ν | lass redu | ictions at: 30 m 150 m |
| Paterson et al. 1980 | Milking center waste and barnyard runoff from 70 cow dairy studied for a 5-yr period | 5 yr | Natural rainfall | Yes | 36 | | 3.4 | Tall fescue | | Silt loam |
| | | | Snow melt Perched water table | | | | | | | |
| Schellinger and Clausen 1992 | Runoff from paved dairy lot to deten- tion pond then VTA subject to natural rainfall | 18 mo | Dairy barnyard | Yes | 22.9 | 0.27 | 2 | Fescue, blu and ryegras | egrass, s mix | |
| Williamson 1999 | Describes and compares design and performance of four VTAs in Kansas for feedlot | 5 mo 5/98 | 350 head beef feedlot | Yes | 239 | 0.23 | 1.2 | Bromegrass | 5 | Sandy loam |
| *Same study, different VTA design | location and | 11/98 for all sites | 300 head beef feedlot | Yes | 427 | 0.97 | 0.75 | Bromegrass | 5 | Silty clay loam |
| *Same study, different VTA | location | 11/98 for all sites | 300 head beef feedlot | Yes | 213 | 0.36 | 2 | Fescue | | Silt loam |
| *Same study, different VTA | location | 11/98 for all sites | 200 head beef feedlot | Yes | 137 | 0.59 | 0.6 | Bromegrass | 5 | Loam |
| Woodbury et al. 2002; Woodbury et al. 2003a; Woodbury et al. 2003b | Settling basin and VTA collects open lot runoff from beef cattle facility | 1997–2003 | 600 head beef feed- lot | Yes | 200 | 3 | 0.5 | Brome gras | s | |

(feedlot drainage area)

**m = reductions calcuated on a mass basis c = reductions calculated on a concentrated basis

Table 9-8 Summary of VTA performance when placed on commercial or research livestock facilities—continued

| Perc | rcent reduction | | | | | | | | | | | | | |
|--------------|---------------------|---------------------------|----------------------|------------|-----------|--------------------|--------------------|--|----------------|-------------------|--------------|----|----------|--------|
| TS | TSS | BOD5 | COD | Total N | TKN | NH ₄ -N | NH ₃ -N | NO ₃ -N | Total P | Ortho-P | FC | FS | E. Coli | ** |
| 1.5 c | rm rair | ı ıfall on 5/ | 14/96 | 1 | 85 | 62 | | S | 5uspended P | 25 | 60 | | | |
| 9.1, 7/27 | 3.6, an /96, 6/2 | d 0.6 cm 1 2/96, and 0 | rainfalls 6/27/98 | on | 35–75 | 35–80 | | | 25-75 | 15–75 | 20 to 80% | | | |
| 65 | | | | 80 | | | 81 | -87 | 77 | | | | | с |
| 71 | | | | 85 | | | 83 | $\begin{bmatrix} -43\\ 86 \end{bmatrix}$ | 83 95 | | | | | c m |
| 93 | | | | 97 | | | 98 | | | | | | | |
| | | | | | | Most m | nass flow re | eduction oc | curred in ir | nfiltration basir | n | | | |
| | 93 95 | TDS 74 68 | | 77 81 | | | | | 84 79 | | 84 85 | | 91 90 | m m |
| | 71 | 42 | | | | 38 | | increase | 7 | | | | | с |
| | | | | | | | | | | | | | | |
| | 84 | 77 99 | | | | 78 97 | | 40 increase | 32 98 | | | | | c c |
| _ | 33 | - | - | | 18 | 15 | | | 12 | 6 | | - | | m |
| | | - | _ | 61.5 | | | | | 28.6 | _ | 78.9 | - | 79.3 | с |
| | | - | _ | 63.7 | | | | | 56.8 | _ | 76.5 | - | 78.2 | с |
| | | - | _ | 19 | | | - | | 13 | - | 36 | 83 | | с |
| | | - | - | 52.8 | | | | - | 74.2 | - | 90.3 | - | 88.4 | с |
| No o | bserv | ed discha | rge of wa | ater below | root zone | e for 2 yr o | or as surfac | ce water fro | m VTA for | 5 yr | | | | m |

Table 9–9Summary of VTA performance under simulated conditions

| Study description | | | VTA info | ormatio | n | | |
|-----------------------------------|----------------------------------|-----------------|---------------|---------|---------|--------------------------|-------------------|
| | | . . | Length | | Slope | | a |
| Reference Cormo at al. 1008 | Summary Four VTA plots placed | 64 mm/b | (m) | AK 1/ | (%) | Tall forgets and | Soll Silt loom |
| Coyne et al. 1998 | after poultry manure | 04 1111/11 | 4.0 | 0.20 | 9 | Kentucky blue- | Siit Ioani |
| | amended pasture area | | | | | grass | |
| | | 64 mm/h | 9 | 0.66 | 9 | Fescue-bluegrass | Silt loam |
| | | | | | | mix | |
| Chaubey et al. 1994 | Swine manure applied | 50 mm/h | 3 | 1 | 3 | Fescue | Silt loam |
| | to VTA subject to simu- | | | | | | |
| | lated rainfall | 50 mm/h | 6 | 2 | 3 | Fescue | Silt loam |
| | | 50 mm/h | 9 | 3 | 3 | Fescue | Silt loam |
| | | 50 mm/h | 15 | 5 | 3 | Fescue | Silt loam |
| | | 50 mm/h | 21 | 7 | 3 | Fescue | Silt loam |
| Chaubey et al. 1995 | Poultry manure | 50 mm/h | 3 | 1 | 3 | Fescue | Silt loam |
| | applied | 50 mm/h | 6 | 2 | 3 | Fescue | Silt loam |
| | to VIA subject to | | | | | | |
| | | 50 mm/h | 9 | 3 | 3 | Fescue | Silt loam |
| | | 50 mm/h | 15 | 5 | 3 | Fescue | Silt loam |
| | | 50 mm/h | 21 | 7 | 3 | Fescue | Silt loam |
| Dillaha et al. 1988: | Simulated feedlot and | 50 mm/h | 4.6 | 0.25 | 11 | Orchardgrass | Silt loam |
| Dillaha et al. 1986 | rainfall | 50 mm/h | 9.1 | 0.50 | 11 | Orchardgrass | Silt loam |
| | | 50 mm/h | 4.6 | 0.25 | 16 | Orchardgrass | Silt loam |
| | | 50 mm/h | 91 | 0.50 | 16 | Orchardgrass | Silt loam |
| | *concentrated flow | 50 mm/h | 4.6 | 0.25 | 5 | Orchardgrass | Silt loam |
| | *concentrated flow | 50 mm/h | 91 | 0.50 | 5 | Orchardgrass | Silt loam |
| Edwards et al. 1983 | VTA test plots after | 50 1111/11 | 2×30 | 0.00 | 2 | Fescue | Silt loam |
| Luwards et al. 1999 | settling basin, natural | | | | | rescue | Shit Iouni |
| | rainfall, | | | | | | |
| | 56 head of beef cattle | | | | | | |
| | on concrete lot | | | | | | |
| Fajardo et al. 2001 | Plot study compar- | 17 mm/h for | 30 | | 4.3-5.1 | Tall fescue | Fine silt |
| | ing fallow vs. vegetat- | 110 mm/h for | | | | | |
| | strip | VTA | | | | | |
| Goel et al. 2004 | A dairy slurry and wa- | 1.2 L/s ap- | 5 | Width | 3 | Perennial rye | Guelph |
| | ter mix was applied | plied to up- | 10 | = | | | loam |
| | to upper end of three | per end of fil- | 5 | 1.2 m | | Mixed grass spe- | |
| | lengths of VTA and | ter strip | 10 | | | cies | |
| | three vegetative covers | |) 1 | | | drass | |
| | were tested | | | | | 51455 | |
| | | | 5 | | | Perennial rye | |
| | | | 10 | | | | |
| | | | 5 | | | Mixed grass | |
| | | | 10 | | | Species Kentucky blue | |
| | | | 10 | | | grass | |

Table 9–9 Summary of VTA performance under simulated conditions—Continued

| Perce | ent rea | uction | <u> </u> | | 1 | | | | | 1 | | . <u> </u> | 1 | 1 |
|-------|----------|--------|----------|--------------|----------|----------|---------|----------|------------|----------|--------|------------|--------|-------------|
| тя | TSS | BOD5 | COD | Total N | TKN | NHN | NHN | NON | Total P | Ortho-P | FC 2/ | FS 2/ | E Coli | |
| 96 | 100 | DODS | | | | 11114-11 | 1113-11 | 1103-11 | - | | 75 | 68 | | c <u>3/</u> |
| 50 | | | | | | | | | | | 10 | | | |
| | | | | | | | | | | | | | | |
| 98 | | | | | | | | | | | 91 | 74 | | c |
| | | | | | | | | | | | | | | |
| | | | | | 65 | | 71 | | 67 | 65 | | | | m 3/ |
| | | | | | | | | | | | | | | pt |
| | | | | | 69 | | 83 | | 71 | 71 | | | | m |
| | | | | | 89 | | 96 | | 87 | 89 | | | | m |
| | | | | | 86 | | 99 | | 91 | 93 | | | | m |
| | | | | | 87 | | 99 | | 92 | 94 | | | | m |
| | | | | | 39 | | 47 | | 40 | 39 | | | | m |
| | | | | | 54 | | 70 | | 58 | 55 | | | | m |
| | | | | | | | | | | | | | | |
| | | | | | 67 | | 79 | | 74 | 71 | | | | |
| | | | | | 76 | | 04 | | 07 | 85 | | | | |
| | | | | | 01 | | 94 | | 01 | 00 | | | | |
| | 97 | | | 61 | 64 | 94 | 90 | 26 | 62 | 20 | | | | |
| | 07 | | | | 04 04 | 60 | | -50 | 80 | 20 | | | | |
| | 90 76 | | | 67 | 60 | 09 | | 4 | 50 | 100 | | | | C |
| | 00 | | | 71 | 09 | -21 | | 0 17 | 57 | 51 | | | | C |
| | 00 | | | | | -30 | | 17 | 01 | -51 | | | | C |
| | 51 | | | | | | | -84 | | -ð 91 | | | | C |
| 07 | 98 | 01 | 00 | (| 9 | -11 | | -198 | 19 | 51 | | | | c |
| 87 | | 81 | 89 | 83 | | | | | 84 | | | | | m |
| | | | | | | | | | | | | | | |
| | | | | | | | | | | | | | | |
| | | | | | | | | | | | | | | |
| | | | | | | | 94–99 | | | | No | | | c |
| | | | | | | | | | | | change | | | |
| | | | | | | | | | | | | | | |
| | 00 | | | 01 | | | | | 00 | 50 | 61 | | CC | |
| | 80 86 | | | 91 | | | | - 45 | 88 | | 53 | | 36 | C |
| | 87 | | | 87 | | | | 25 | 87 | 44 | 15 | | -26 | |
| | 91 | | | 84 | | | | 16 | 86 | 48 | 52 | | 58 | |
| | 89 | | | 92 | | | | 13 | 89 | 50 | 68 | | -130 | C |
| | 91 | | | 95 | | | | 35 | 92 | 58 | 74 | | 77 | C |
| | | | | | | | | | | | | | | C |
| | 90 | | | 94 | | | | 3 | 91 | 64 | 71 | | 67 | m |
| | 94 | | | 95 | | | | 67 | 95 | | | | 64 | m |
| | 91 | | | 89 | | | | 49 | 90 | 66 | 56 | | 58 | m |
| | 95 | | | 91 | | | | 92 75 | 92 | 10 | (5) | | 82 | |
| | 97 | | | 98 | | | | 10 | 9/ | 80 | 91 | | 39 | 111 |
| | 99 | | 1 | 100 | | | | 90 | 100 | 91 | 99 | | 99 | Im |

Percent reduction

Table 9-9 Summary of VTA performance under simulated conditions—Continued

| Study description | | | VTA information | | | | | |
|--|--|--|----------------------------|---------------------------|--|--|--|--|
| Reference | Summary | Intensity | Length (m) | AR* | Slope (%) | Vegetation | Soil | |
| Hawkins et al. 1998 | WW pumped from swine lagoon to VTA; runoff and percolate | | 6.1 | | 5 11 | Bermuda and ryegrass mix | Loamy sand | |
| | analyzed | | | | | | | |
| Lim et al. 1997 | Simulated pasture and | 10 cm/h | 6.1 | 0.50 | 3 | Fescue | Silt loam | |
| | | 10 cm/h | 12.2 | 1.00 | 3 | Fescue | Silt loam | |
| Prantner et al. 2001 | Lab scale study of raw swine manure applied to soil infiltration areas | | 10.3 | 1.0 | 3 | None | Clarion loam soil | |
| Sanderson et al. 2001 | Manure application to grassland with VTA down gradient | | 16.4 | 1.0 | 1 | Switchgrass | Fine sandy loam | |
| Schwer and Clausen, 1989 | VTA test plot, natu- ral rainfall, milk house waste water pumped to VTA | | 26 | | 2 | Fescue, ryegrass, bluegrass mix | Sandy loam | |
| | Same VTA, subsurface flow analysis | High rate: 20–27 cm/wk Low rate: 6–16 cm/wk | | | | Fescue low rate high rate | Loam (surface); clay loam (sub- surface) | |
| Srivastava et al. 1996 | Nine control VTA plots placed after manure amended pasture | | 3.1–18.3 | 3 1 0.33 | 3 | Fescue | Silt loam | |
| Young et al. 1980 | Rainfall simulator ap- plied 25-yr, 24-h storm to VTA plots containing corn, orchardgrass, sor- ghum-Sudangrass mix over 2-yr test period | 6.35 cm/h for 71 min | 27 27 27 21 21 | 2 2 2 1.6 1.6 | $\begin{array}{c}4\\4\\4\\4\\4\\4\\4\end{array}$ | Corn Orchardgrass Sorghum- Sudangrass mix Corn Oats | Runoff volume reduction 98% 81% 61% 66% 41% | |
| Willrich and Boda, 1976 | VTA test plots, natural rainfall, swine lagoon effluent pumped to VTA | | 30.5 | | 3 | Fescue | Clay loam | |
| | | | 6.1 | | 5 | Bermuda and ryegrass mix | Loamy sand | |
| *Same source of wastew with different slope | vater pumped to VTA | | 6.1 | | 11 | Bermuda and ryegrass mix | Sandy loam | |
| | | | 6.1 | | 11 | Bermuda and rvegrass mix | Sandy loam | |

1/ AR = area ratio = VTA area/feedlot drainage areas

2/ FC = fecal coliform; FS = fecal streptococci

3/c = reductions calculated on a concentration basis; m = reductions calculated on a mass basis

4/ Data represents total organic carbon as measured by Srivastava et al. 1996

Table 9-9 Summary of VTA performance under simulated conditions—Continued

| $ \begin{array}{c ccccccccccccccccccccccccccccccccccc$ | Percent reduct | ion | | | | - | | | ~ | | - | | | | |
|---|----------------|----------|------------------|-----------|------------|----------|--------------------|--------------------|--------------------|----------|---------|-----|----|--------------------|----|
| $\begin{array}{c ccccccccccccccccccccccccccccccccccc$ | TS | TSS | BOD ₅ | COD | Total N | TKN | NH ₄ -N | NH ₃ -N | NO ₃ -N | Total P | Ortho-P | FC | FS | E. Coli | ** |
| $\begin{array}{c ccccccccccccccccccccccccccccccccccc$ | 14 | | | 52 | | 3 | | 1 | 47 | 22 | | | | | с |
| $\begin{array}{c ccccccccccccccccccccccccccccccccccc$ | 5 | | | 81 | | 60 | | 58 | 54 | 75 | | | | | m |
| $\begin{array}{c c c c c c c c c c c c c c c c c c c $ | -557 | | | 14 | | 33 | | 33 | -834 | -11 | | | | | с |
| $\begin{array}{c ccccccccccccccccccccccccccccccccccc$ | 37 | | | 92 | | 93 | ļ | 93 | -59 | 92 | | | | | m |
| $ \begin{array}{c ccccccccccccccccccccccccccccccccccc$ | 23.6 | 70 | | | | 78 | | 18.6 | -498.2 | 76.1 | 74.5 | 100 | | | m |
| $ \begin{array}{c c c c c c c c c c c c c c c c c c c $ | 40.8 | 89.5 | | | | 89.5 | | 52.8 | -140.1 | 90.1 | 87.8 | 100 | | | m |
| $\begin{array}{c c c c c c c c c c c c c c c c c c c $ | 69.8 | 97.6 | | | | 95.3 | | 68 | -96.7 | 93.6 | 93 | 100 | | | m |
| $\begin{array}{c c c c c c c c c c c c c c c c c c c $ | 1998 Ur | ndiluted | swine | manur | e | 94 | | | Inc. | 85 | | | | | с |
| 1999 Undiluted swine manure 96 1-77 98 c c c 3 parts manure + 1 part water 94 mg/L 97 98 1 c c 92 25- 44 44 83 46 86 82 c c 97 93 70 92 62 94 13 c c 97 93 70 92 90 c c 97 93 70 92 90 c c 1urbidity 31 15 26 14 13 c 1urbidity 31 15 26 26 14 13 c 11 30 67 75 66 62 66 </td <td>3 par</td> <td>ts man</td> <td>ure + 1</td> <td>part w</td> <td>ater</td> <td>87</td> <td></td> <td></td> <td>from</td> <td>78</td> <td></td> <td></td> <td></td> <td></td> <td>с</td> | 3 par | ts man | ure + 1 | part w | ater | 87 | | | from | 78 | | | | | с |
| $\begin{array}{c c c c c c c c c c c c c c c c c c c $ | 1999 Ur | ndiluted | swine | manur | e | 96 | | | 1–77 | 98 | | | | | с |
| $\begin{array}{c c c c c c c c c c c c c c c c c c c $ | 3 par | rts man | ure $+1$ | part w | ater | 94 | | | mg/L | 97 | | | | | с |
| $\begin{array}{c c c c c c c c c c c c c c c c c c c $ | | | | 25– 44 | | | | | | | 4-76 | | | | m |
| $\begin{array}{c c c c c c c c c c c c c c c c c c c $ | | 92 | | | | 83 | 46 | | | 86 | 82 | | | | с |
| Turbidity 31 31 15 67 50 26 62 43 14 62 46 13 c n n 31 15 67 50 62 43 62 43 14 46 13 13 c n n 30 67 11 $13^{4'}$ 75 21 66 39 27 36 26 14 13 a | | 97 | | | | 93 | 70 | | | 92 | 90 | | | | с |
| $\begin{array}{c c c c c c c c c c c c c c c c c c c $ | Turbidity | | | | | | | | | | | | | | |
| $\begin{array}{c c c c c c c c c c c c c c c c c c c $ | 31 | | 31 | 15 | | | | 26 | | 14 | | 13 | | | с |
| $\begin{array}{c c c c c c c c c c c c c c c c c c c $ | | | | 67 | | | 62 | | | 62 | | | | | m |
| $\begin{array}{c c c c c c c c c c c c c c c c c c c $ | | | | 50 | | | 43 | | | 46 | | | | | m |
| $ \begin{array}{ c c c c c c c c c c c c c c c c c c c$ | | | | 30 | | 67 | | 75 | | 66 | | | | | |
| $13^{4'}$ 21 27 26 | | | | 11 | | 44 | | 39 | | 36 | | | | | |
| | | | | 134/ | | 21 | | 27 | | 26 | | | | | |
| Total coliforms | | | | | | | | | | | | | | Total coliforms | |
| Sediment | Sediment | | | | | | | | | | | | | | |
| 93 98 98 95 98 100 53 | 93 | | | | | 98 | 98 | | 95 | 98 | 100 | | | 53 | |
| $\begin{array}{c ccccccccccccccccccccccccccccccccccc$ | 66 | | | | | 69 | 65 | | 9 | 76 | 77 | 55 | 72 | 81 | |
| 82 50 47 -81 48 42 83 68 | 82 | | | | | 50 | 47 | | -81 | 48 | 42 | 83 | 68 | | |
| | 01 | | | | | 70 | 70 | | 441 | 74 | 41 | | | 71 | |
| 75 75 75 75 75 75 75 75 | 01 75 | | | | | 79 45 | 10 | | -441 | 74 50 | 41 | | | 70 | |
| | 15 | | | | | 40 | 55 | | -1150 | 50 | -0 | | | | |
| 31 15 26 14 31 | | | 31 | 15 | | | | 26 | | | 14 | 31 | | | |
| | | | | | | | | | | | | | | | |
| 59 81 60 58 54 75 | 59 | | | 81 | | 60 | 58 | | 54 | 75 | | | | | |
| -557 14 33 33 -834 -11 | -557 | | | 14 | | 33 | 33 | | -834 | -11 | | | | | |
| 37 92 93 93 -59 92 | 37 | | | 92 | | 93 | 93 | | -59 | 92 | | | | | |

| Reference | Type of system |
|---|--|
| Barker and Young 1984 | Milking center wastewater and open lot runoff from a 54 cow dairy was directed to settling basin and VTA. Four earthen berms located at 30-in intervals were designed to create a cascading type system. System was monitored over 2 yr |
| Coyne et al. 1998 | Controlled replicated research trials were conducted on VTA of 4.5 m and 9.0 m in length below a simulated pasture area with poultry manure added. A 64-mm/h rainfall was applied |
| Chaubey et al. 1995 | Poultry manure applied to established grass area with VTA located below area of land application. Site is subject to simulated rainfall |
| Dickey and Vanderholm 1981a; Vanderholm and Dickey 1980; Dickey and Vanderholm 1981b | Papers review design and performance of four VTA, two functioning as overland flow (100 cow dairy and 450 beef feedlot) and additional two as channelized flow (500 head beef feedlot and 480 swine operation) |
| Dillaha et al. 1988; Dillaha et al. 1986 | Controlled replicated research trials were conducted on VTA of 4.6 m and 9.1 m in length below a simulated dairy open lot of 18.3 m on a silt loam soil. A 50-mm/h rainfall was applied for 2 h on soils described as "dry," "wet," and "very wet" |
| Edwards et al. 1983 | VTA test plots after settling basin, natural rainfall, 56 head of beef cattle on concrete lot. Two grass filter cells were used in series, each representing approximately 50% of the concrete lot area |
| Edwards et al. 1986 | VIB used with 56 head of beef cattle on concrete lot. VIB was preceded by solids settling basin |
| Fausey et al. 1988 | |
| Fajardo et al. 2001 | VTA and fallow plots are placed below area of manure application. Sufficient simulated rainfall was applied to achieve 1-h runoff event. Much greater volumes were applied to VTA plots |
| Harner and Kalita 1999 | VTA established on several open lot beef systems in three watersheds, three of which were moni- tored for performance |

Table 9–10Summary of performance observations for VTA for past research and field demonstration projects—Continued

Performance observations (in addition to % reductions reported in tables 9-8 and 9-9)

- Effluent leaving the VTA effluent was only 5% of VTA influent volume resulting is high pollutant mass reductions
- Increased soil nitrates were observed in deep soil samples in sections prior to first two berms. Increased soil P levels were also observed ahead of first two berms. No other soil samples showed increases
- Soluble salt concentration showed increases in all soil samples ahead of first two berms. Total cations remained relatively constant with exception of shallow soil samples taken ahead of first berm
- VTA distribution pipe at upper end of field with four separate outlets produced channel flow concerns. Increasing number of outlets to seven appeared to reduce channel flow concerns
- 85% and 76% of total water runoff infiltrated into the 9.0 m and 4.5 m VFA plots, respectively
- The 4.5-m VTA trapped most of the sediment in runoff
- VTA of this length trapped most of the fecal bacteria that moved onto the site. However, the concentration of fecal bacteria in runoff remained high and exceeded water quality standards
- · First order linear regression describes reduction in mass transport of litter constituents with VTA length
- Removal of contaminants in VTA increased for lengths up to 15.2 m (ammonia and dissolved phosphorus), 9.2 m (total Kjeldahl nitrogen and total phosphorus), and 3.1 m (total suspended solids and chemical oxygen demand)
- VTA holds promise for improving quality of runoff from land application sites treated with poultry litter
- VTA reduces nutrients, solids and organic matter from feedlot runoff by more than 80% on a concentration basis and 95% on a weight basis
- Additional removals are impractical due to quality of runoff approaching that of agricultural land that is not exposed to feedlot runoff. Discharge did not meet stream quality standards
- Fecal coliform levels from the VTA with feedlot runoff addition were one log higher than runoff from a control VTA with no manure addition. Both were high in relation to stream standards
- Most runoff events infiltrated completely, resulting in no discharge. Sizing procedures used for project resulted in runoff only during large precipitation events and high stream flows
- VTA are effective for removal of sediment and suspended solids with filters of 9.1 m or less if flow is shallow and uniform
- Some decline in effectiveness is noted with time as sediment accumulates
- Total N and P are not removed as effectively as sediment for the lengths tested
- VTA lengths used in this research were not effective in removing soluble N and P. Soluble P was often higher in outflow than inflow, presumably due to release of P previously trapped in the VTA
- VTA with concentrated flow were significantly less effective than were uniform flow plots
- Settling basin and filter strips reduced contaminant mass transport by 81–89%
- The settling basin was more effective in large storm events
- The grass filter strip was more effective when the basin was slowly drained 1 day following a storm event
- Infiltration basin approach eliminated all overland flow runoff to receiving stream
- Infiltration basin produced greater nutrient transport reduction than a 33-m grass filter strip but was less effective than a 66-m grass filter strip
- Reed canarygrass thrived in the infiltration basin
- Drain tile placed across the slope in the infiltration basin produced greater discharge volumes and greater pollutant transport from the drain tiles than a single drain tile placed parallel with the slope of the infiltration basin
- Bacterial contamination in runoff water was not reduced when comparing tall fescue and fallow filter strips. Presence of bacterial organisms on the soil surface is ubiquitous. Manure addition did not significantly impact source of bacterial organisms
- Dilution due to substantially greater water application in VTA to achieve similar runoff many also be partial explanation for reduced nitrates and unchanged coliform concentration (Author's note: all comparisons are based only on concentration)
- VTA effectively reduces nuitrient, sediment, and bacteria from open lot livestock systems
- Quality of vegetation impacts nutrient uptake capacity of VTA

| Reference | Type of system |
|------------------------|--|
| Hawkins et al. 1998 | Effluent pumped from swine lagoon to VTA; runoff and percolate analyzed |
| Hubbard et al. 1994 | Pre-treated swine lagoon effluent was applied at a rate of 450 and 900kg/ha/yr to three VTA consist- ing of 1) 10-m wide grass (bermuda and tall fescue) followed by 20-m riparian zones, 2)10-m grass and 20-m maidencane zones and 3) 20-m grass and 10-m riparian zones |
| Hubbard et al. 1999 | Pre-treated swine lagoon effluent was applied at a rate of 800 kg N and 150kg P per ha per yr to six different wetland and riparian plant species to evaluate plant response. |
| Komor and Hansen 2003 | A settling basin and VTA is applied to two small feedlot sites in Minnesota (200 and 225 cattle capacity). Data was collected for seven rainfall events ranging from early May to late October. VTAs were sized to represent approximately 20% of the feedlot runoff area |
| Lim et al. 1998 | Cattle manure was applied to upper 12.2 m of grassed plots. Runoff was collected at 0, 6.1, 12.2, and 18.3 m below area of manure application for simulated rainfall of 100 mm/h |
| Lorimor et al. 2003 | Runoff from 380 head concrete feedlot passes through settling channel (stage 1), infiltration basin (stage 2), and wetlands (stage 3). |
| Mankin and Okoren 2003 | 300 head heifer feedlot with runoff directed to settling basin (1st stage) and VTA (stage 2) |
| Nienaber et al. 1974 | Settling basin, holding pond, sprinkler irrigation on grassed treatment area. Fresh water application compared with beef feedlot runoff |
| Paterson et al. 1980 | Milking center waste and barnyard runoff from 70 cow dairy was directed through settling basin (stage 1), holding tank with lift pump, and VTA (stage 2). |
| Prantner et al. 2001 | Undiluted swine manure, 3 to 1 swine manure and water, and water applied to buried containers with grass (first stage) followed by wetland plants (stage 2). Sufficient manure or water volume applied at 2-wk intervals to saturate soil column |

Table 9-10Summary of performance observations for VTA for past research and field demonstration projects—Continued

Performance observations (in addition to % reductions reported in tables 9-8 and 9-9).

- Significant nitrification occurred on the steeper slope and elevated soil nitrate levels were a concern
- Intense monitoring of nitrogen in soil, ground water, and surface water runoff was reported for a 9-month period with no differences in treatments observed at this time
- All three treatments were effectively filtering N from applied swine manure at both rates
- Significant reductions in ammonium in surface runoff were noted with down gradient distance from point of swine manure application. Nitrate concentration increased from less than 1 mg/L to between 1 and 15 mg/L
- All species responded well to swine effluent application with buttonbush and saltmeadow cordgrass showing the greatest growth response
- Significant variation occurred in performance of VTA for different rainfall events. Greatest attenuation occurred on October and May when mats of wilted, flat-lying grass covered the filter strips. Attenuation was least during the summer when tall growing grass covered the filter strips
- On one site, runoff volume was reduced from 47% for a 2.3-cm (spring rainfall) to 100% for a fall 1.5-cm fall rainfall event. On the second site, runoff volumes were reduced by 83% for a 3.6-cm fall event, 85% for a 9.1-cm summer event, to 98% for 0.6-cm summer rainfall event
- Ground water degradation was observed where shallow water table exists (1.3 m and 0.8 m below ground surface at two sites)
- No concentration reductions were observed after first 6.1 m
- Concentration and mass transport reductions of the analyzed parameters followed a first-order exponential reduction relationship with length of VTA
- Overall mass flow reductions have been between 86 and 98% for this system, with most significant reductions due to VIB
- After 5-yr use, soil phosphorus levels within the infiltration basin have not shown signs of buildup
- Although the flow out of the infiltration basin is not continuous, it has a substantially lower peak and extended period of flow as compared to the runoff flow from the feedlot. The infiltration basin also stores significant quantities of water subsequently used by plant growth thus reducing total volume. This change in flow pattern is beneficial to secondary treatment systems
- Mass reduction of constituents occurred in first 30 m. Little or no reduction occurred in last 120 m
- Fecal coliform concentration was reduced below accepted water quality standards
- Application rates of 64 cm (25 in) in 1971 and 91 cm (36 in) in 1972 did not result in runoff (applied mid spring through late fall) or accumulation of nitrogen, phosphorus, or chlorides
- Four pollutants (BOD, NH4, PO4, and suspended solids) decreased in concentration by passing though VTA
- Four pollutants were reduced by 97% or more in perched ground water while nitrate increased
- Nitrate increased during passage through VTA except during winter where nitrate was reduced in concentration
- Systems were designed to encourage nitrification followed by denitrification processes and soil absorption and settling of phosphorus. The 2-yr study produced 99.5% and 99.9% reduction in ammonium-N, 98.5% and 99.8% reduction in total P and ending nitrate concentrations of 0.2 mg/L (1998) and 7–9 mg/L (1999). Similar percentage of reduction of ammonium and phosphorus were observed in the infiltration and wetland zones. Soil P accumulation was a concern but not observed in the 2-yr study

Table 9–10

| Reference | Type of system |
|---------------------------------|---|
| Sanderson et al. 2001 | Solid dairy manure (1995) and dairy lagoon effluent (1996 and 1997) was applied at rates ranging from 0–600 kg N/ha in a replicate plot design. Manure was applied to a switchgrass area with a VTA consisting of switchgrass below the manured plots |
| Scheilinger and Clausen 1992 | Concrete dairy barny ard runoff flows through a detention pond and into a $22.9~{\rm m}$ by $7.6~{\rm m}$ VTA with 2% slope |
| Schmitt et al. 1999 | Alternative lengths of VTA and types of vegetation were evaluated for agricultural field runoff |

Summary of performance observations for VTA for past research and field demonstration projects-Continued

| Schwer and Clausen 1989 | VTA was designed to treat milk house wastewater on a Vermont dairy |
|----------------------------|---|
| Srivastava et al. 1996 | Nine control VTA plots, ranging from 3–18.3 m, were placed after poultry manure amended pasture |

Willrich and Boda 1976 Anaerobic lagoon swine effluent is applied to upper end of six plots

| Woodbury et al. 2002; | Runoff from eight open lot beef cattle pens (about 600 cattle) moved from the pens through a grass |
|------------------------|--|
| Woodbury et al. 2003a; | approach, settling basin (created by a 300-m long terrace below the pens), and a 6-ha VTA |
| Woodbury et al. 2003b | |
| | |

| Young et al. 1980 | Rainfall simulator applied 25-yr, 24-h storm to VTA plots containing corn, orchardgrass, sorghum- Sudangrass mix, oats over a 2-yr test period |
|--------------------|---|
| Younos et al. 1998 | 18-m wide VTA placed down gradient from open lot for 60 head dairy |

Table 9-10 Summary of performance observations for VTA for past research and field demonstration projects—Continued

Performance observations (In addition to % reductions reported in tables 9-8 and 9-9)

- VTA effectively reduced total reactive P and COD concentrations in surface runoff
- Runoff concentration of N, P, and COD decreased as greater time lapsed between manure application and precipitation event. To minimize N and COD runoff concentrations, 3–4 days was suggested. To minimize P concentrations, then 1 day was necessary
- 65% of barnyard runoff exited from VTA. Retention of solids, N, P, K, and bacteria was considered poor
- Average hydraulic retention time of 15 min was observed
- Inadequate detention time and excessive hydraulic detention times were identified as reasons for poor performance
- VTA performance is strongly dependent upon type of contaminants. VTA are most effective for sediment related contaminants and least effective for dissolved contaminants
- Doubling filter strip from 7.5–15 m does not improve sediment settling, increases infiltration, and increases dilution of runoff
- Incorporating trees and shrubs into the lower half of filter strips does not affect performance
- Contour sorghum strips of equal width are not as effective at reducing contaminants as perennial vegetation
- Retention was greatest during the growing season and least during snowmelt
- Retention of N and P in harvested crops accounted represented only a small portion of input nutrients
- Pollutant concentration of water exiting litter treated areas is not dependent on litter treated length, suggested rapid equilibrium being reached
- · Pollutant concentrations decreased with increasing VTA length for all pollutants
- Mass transport was not affected by VTA length with large portion of the mass removal occurring within the first 3 m of VTA
- Overland flow treatment of swine lagoon effluent caused significant concentration attenuations and mass reductions of its polluting properties
- BOD and turbidity removal became effective with time whereas treatment effectiveness for COD, phosphorus, salinity and ammonia decreased with time
- · Changes in application rate impacted runoff volumes but did not significantly change concentration of most contaminants
- Significantly greater attenuation occurred during cool, wet months for turbidity and fecal coliform and during warm, dry months for phosphorus. Nitrification was also greater during warn, dry months.
- The settling basin removed 80,67, 59, and 47% of the total suspended solids, volatile suspended solids, chemical oxygen demand, and total nitrogen
- Distribution of settling basin water to a VTA was not uniform resulting in soil nitrate accumulation in upper 30 cm (1 ft)
- No water was measured exiting the VTA below the root zone or at the down gradient end of the VTA over a 3-yr period suggesting hay crop utilization of all applied water
- · Mass nitrogen removal by harvesting exceeded mass nitrogen addition with feedlot runoff
- Migration of nitrate below the settling basin is a problem, possibly exacerbated by solids removal and basin cleaning
- Significant reductions on nitrogen forms (with exception of nitrate), phosphorus, and microorganisms were observed for 36 m VTA
- Nonstructural control practices are a promising alternative method for controlling feedlot runoff
- Stream loads for total runoff, orthophosphate and dissolved phosphorus, total phosphorus, and total nitrogen were lower after VTA installation as compared to a pre-VTA installation. However, due to the relatively short monitoring (6 mo prior and after installation), differences were statistically inconclusive
- Although the water quality upstream of the sacrifice lot is already degraded, the installation of the VTA may prevent a further degradation of the water quality downstream of the sacrifice lot

Nitrogen removal—The most common gauges of nitrogen content in surface runoff include total nitrogen (TN), total Kjeldahl nitrogen (TKN), ammonium and ammonia nitrogen (NH₄ and NH₃, respectively), and Nitrate (NO₃) (Ikenberry and Mankin 2000). Removal of TN, TKN, NH₄, and NH₃ by VTA, has been shown to exceed 85 percent. Nitrate (NO₃) removal has typically been much lower, although Fajardo et al. (2001) reported 97 and 99 percent reductions in simulated VTA studies. In some studies, NO₃ increased from nearzero levels typical of most anaerobic feedlot runoff, to sub-health-limit levels during flow through the VTA. Chaubey et al. (1995) noted that removal of ammonia and TKN in VTA increased for lengths up to 15.2 and 9.2 meters, respectively. Overall properly designed and managed VTAs are very effective, averaging approximately 70 percent nitrogen removal (Ikenberry and Mankin 2000).

Phosphorous removal—Because the majority of the phosphorous in feedlot runoff is adsorbed to solids particles, total phosphorous removal is directly related to solids removal efficiencies. Phosphorous removal rates have ranged from 12 to 97 percent, averaging about 70 percent. Chaubey et al. (1995) also noted that removal of dissolved and total phosphorus in VTA increased for lengths up to 15.2 meters and 9.2 meters, respectively.

Vegetative infiltration basin (VIB)

Some vegetative systems force infiltration of runoff through a soil filter and provide an alternative approach that prevents surface water discharges. Lorimor et al. (2003) operated a bermed infiltration area that allowed discharges only through subsurface drain tiles placed 1.8 meters (6 ft) below the surface of this basin. All runoff must move through a soil filter prior to discharge. Smaller footprint for the VTA (1/6 to 1/12 of most standard VTA designs) and no direct surface-water discharge are two advantages. After 5 years of experience, soil P levels have not shown signs of buildup. Preferential flow through the soil filter may be a potential concern over time. Infiltration basins represent an alternative VTA design that out-performs most grass filters but may be acceptable only for sites with low-infiltration clay layer below the drain-tile. Edwards et al. (1986 and 1988) have reported operation of an infiltration basin below a small open lot cattle facility (table 9-8).

As wastewater infiltrates the soil, aerobic nitrification occurs, converting ammonium to nitrate by the aerobic bacteria *Nitrosomonas* and *Nitrobacter* (Prantner et al. 2001). In addition, phosphorus interacts and becomes attached to soil particles in the profile. Field drainage tile is used to intercept the filtrate and carry it to a secondary form of treatment such as a constructed wetland or VTA.

Two recent infiltration studies at Iowa State University have shown significant water quality improvements. Using liquid swine manure, Prantner et al. (2001) showed over 93 percent reductions in NH₄–N, and 89 percent reduction in phosphorus. Yang and Lorimor (2000) reported a field infiltration system down gradient of a 380-head concrete beef feedlot. Over 2 years of sampling, they found an 81 percent reduction in suspended solids, 83 percent reduction in TKN, 85 percent reduction in NH₄–N, and a 78 percent reduction in P. Nitrate levels have increased by 87 percent, suggesting a need for nitrate utilization or treatment downstream of an infiltration system.

Infiltration basins based upon soil filters are limited to sites conducive to tile drainage where a restrictive soil layer exists below the surface restricting water and contaminant movement to ground water. Alternative infiltration systems, such as a constructed infiltration bed of sand, biosolids, and wood chip mixtures laid over a gravel layer with a tile drain used to treat runoff from paved parking lots (Culbertson and Hutchinson 2004), may have application to livestock systems.

Another advantage of an infiltration basins is its ability to alter the flow rate and timing of liquid (hydrograph) exiting the infiltration basin (Lorimor et al. 2003). Slowing the flow from the infiltration basin during the storm event and delaying much of the discharge until after the storm event enhances the potential for successful treatment in later treatment components such as a VTA.

Overall VTS performance

By coupling various combinations of treatments into a treatment system, the quality of feedlot runoff can be significantly improved to the point of achieving functional equivalency to baseline technologies to complete elimination of surface water runoff. Although the particular combination of treatments selected for any feedlot will be site specific, essentially all should begin with solids settling. Table 9–11 shows a summary of the anticipated contaminant reductions for various treatment components associated with a dairy or beef open lot facility. Reductions for two or more components can be estimated by multiplying remaining contaminants (one reduction) for each component. A settling basin and VIB will reduce total solids concentration by 92 percent or 100-[(100-60) x (100-80)].

VTA design

The literature provided illustrations of a number of critical design considerations for VTAs (table 9–12). Based upon this literature, there are several design considerations that are generally accepted for VTAs:

- A need exists for some degree of pretreatment. Solids settling is commonly used with VTAs to minimize solids accumulation at the front end of a VTA. This pre-treatment minimizes vegetation damage and reduces the potential for channel flow paths and vegetation damage where runoff first enters the VTA.
- Uniform sheet flow of liquid is essential for optimum VTA performance. Design of inlets and headlands is critical to initiating sheet flow. Field management is critical to minimizing concentrated flow. Even with the best inlet design and management, concentrated flow is likely to occur within a VTA and may requiring additional structures to redistribute flow.
- For VTS on CAFOs, minimizing potential for discharge will be critical for achieving equal or better performance than baseline technologies. Combinations of treatment components into systems, attention to sizing, and modification of hydrograph of flow into a VTA are important considerations for minimizing discharge potential.
- Siting criteria is critical to the appropriate application of VTAs. Iowa Department of Natural Resources has established nine evaluation criteria used to initially judge a site including available area, soil permeability, depth to water table,

subsoil and geology, slope, spreaders for uniform distribution, berming for inflow water protection, flooding potential, and proximity to waters of the state (Iowa Department of Natural Resources 2004).

Multiple approaches have been suggested for VTA sizing:

- Dickey and Vanderholm (1981a) recommended a minimum VTA width of 61 meters (200 ft) and a length adequate to completely infiltrate the feedlot runoff and rainfall from a 1-year, 2-hour storm. They calculated minimum flow lengths to provide 2-hour contact times. Based on their model, minimum lengths varied from 91 meters (300 ft) for a 0.5 percent slope up to 262 meters (860 ft) for a 4 percent slope. They also recommended that an infiltration area be designed to allow infiltration for all runoff from a 1-year, 2-hour storm.
- Nienaber et al. (1974) suggested a disposal area of a half hectare per hectare of feed lot is needed. Data in figure 9–4 suggest that a ratio of 1 to 1 (disposal to feedlot area) or greater is necessary to achieve peak performance. Lorimor et al. (2003) has achieved high contaminant removal rates with a ratio of 1 to 6 (infiltration basin to feedlot area) for a bermed infiltration area that allows discharges only through subsurface drain tiles.

| | Total solids | TKN | Ammonium- N | Total P | BOD | |
|----------|--------------|-----|-------------|---------|-----|--|
| Settling | 60 | 80 | 80 | 80 | _ | |
| VTA | 60 | 70 | 70 | 70 | 75 | |
| VIB | 80 | 80 | 85 | 80 | | |
| Wetland | 60 | 50 | 50 | 50 | 60 | |

 Table 9–11
 Summary of contaminant concentration reductions

| Table 9–12 Summary of projects | of design and management recommendations for VTA for past research and field demonstration | | |
|---|---|--|--|
| Reference | Type of system | | |
| Barker and Young 1984 | Milking center wastewater and open lot runoff from a 54 cow dairy was directed to settling ba- sin and VTA. Four earthen berms located at 30 ft intervals were designed to create a cascading type system | | |
| Dickey and Vanderholm 1981a; Vanderholm and Dickey 1980; Dickey and Vanderholm 1981b | Papers review design and performance of four VTA, two functioning as overland flow (100 cow dairy and 450 beef feedlot) and additional two as channelized flow (500 head beef feedlot and 480 swine operation) | | |
| | | | |

| Dillaha et al. 1988; Dillaha, et al. 1986 | |
|--|---|
| Edwards et al. 1983 | • VTA test plots after settling basin, natural rainfall, 56 head of beef cattle on concrete lot. Two grass filter cells were used in series, each representing approximately 50% of the con- crete lot area |
| Ikenberry and Mankin 2000 | Review of literature |

Table 9–12 Summary of design and management recommendations for VTA for past research and field demonstration projects—Continued

| Design recommendations | Management recommendations | | |
|--|---|--|--|
| Initial seeding of fescue and reed canarygrass was used due to tolerance to wet conditions Four distribution points at upper end of VTA proved inadequate to create uniform flow. Later expansion to seven distribution points reduced problems of channel flow | At conclusion of study, orchardgrass and foxtail grass were dominant species at upper end of filter strip and hairy crabgrass dominated in drier areas. Four grass cuttings were made per year with an attempt to hold grass height near 6–12 in high. | | |
| Solids settling in advance of a VTA minimizes vegetation damage and maintains VTA effectiveness Overland or sheet flow within VTA Minimum recommend contact time for runoff with a VTAis 2 h Overland VTA do not require longer contact time as lots increase in size Infiltration area should be designed to allow infiltration for all runoff from a 1-yr, 2-h storm. Additional area provides little improvement Slope and soil infiltration rate are important considerations in VTA sizing Channelized flow systems will: Require flow distances at least 10 times greater that sheet flow design Require one additional hour of contact time beyond the 2-hour minimum for each 465 m² (5,000 ft²) of open lot greater than 929 m² (10,000 ft²) Require large areas for open lots of more than 0.4 ha (1 a) | • Dormant residues in VTA have proven to be an effective filter and settling mechanism. Management practices that contribute to a strong fall growth and well-established dormant residue through winter has value in pollutant removal from winter precipitation and snowmelt runoff | | |
| Effectiveness of VTA is dependent upon design and management measures that create shallow uniform flow and prevent concentrated flow VTA site selection should target flat areas and avoid hilly terrain | See first bullet under design recommendations | | |
| | • The grass filter strip was more effective when basin release was actively managed and slowly drained one day following a storm event and after settling of solids | | |
| | Key management considerations recommended: Soil testing to determine fertilization requirement at time of planting of vegetation Reseeding and fertilization to maintain dense stand Repairing of gullies soon after their development Regular moving and harvesting of plant material to remove nutrients and maintain dense vegetation stand Restriction of field traffic and grazing during wet periods to avoid development of ruts leading to channel flow and damage to vegetation | | |

| Reference | Type of system | | | |
|--|--|--|--|--|
| Lorimor et al. 2003 | Type of system Runoff from concrete open lot beef facility is directed to settling basin, totally bermed in- filtration basin, and constructed wetland | | | |
| Murphy and Bogovich 2001 | Summarizes NRCS design recommendations for application of VTA to open lot dairies in PA for handling runoff and milking center effluent | | | |
| Nienaber et al. 1974 | Settling basin, holding pond, sprinkler irrigation on grassed treatment area. Fresh water application compared with beef feedlot runoff | | | |
| Norman and Edwards 1978 Paterson et al. 1980 | Ohio NRCS recommendations for sizing of buffer strip dimensions for cattle feedlots Milking center waste and barnyard runoff from dairy was directed through settling basin (first stage), holding tank with lift pump, and VTA (second stage) | | | |
| | | | | |
| Murphy and Harner 1999; Harner and Kalita 1999 | VTA established on several open lot beef systems in three watersheds, three of which we monitored for performance | | | |
| Murphy and Harner 2001 | | | | |
| Scheilinger and Clausen 1992 | Runoff from dairy barn yard is directed through a detention pond and then to a VTA | | | |
| Woodbury et al. 2002; Woodbury et al. 2003a; Woodbury et al. 2003b | Runoff from eight open lot beef cattle pens (about 600 cattle) moved from the pens through a grass approach, settling basin (created by a 300-m long terrace below the pens), and a 6 ha VTA | | | |

| Table 9–12 | Summary of design and management recommendations for VTA for past research and field demonstration |
|------------|--|
| | projects—Continued |

| De | sign recommendations | Management recommendations |
|----|---|--|
| • | Infiltration basin was bermed to provide total containment of | Analycinent reconditionations |
| | 25-yr, 24-h storm | |
| • | Infiltration basin was size to provide a land area that was 1/6 | |
| | of the drainage area of the concrete open lot | |
| ٠ | Three parallel buried tile lines ran the length of the infiltration | |
| _ | basin to move filtrate from the basin to a constructed wetland | |
| • | 15 min flow through time for sheet flow at depths of 1.3 cm | |
| | and less for various flow rates and slopes | |
| ٠ | Pretreatment settling basin volume was recommended to be | |
| | 2-yr peak flow times 15 min | |
| V | TA size – Annual feedlot runoff (a-in) | • Applied effluent to a grassed disposal area plant- ad with a mixture of pipe cool and warm season |
| • | Max. annual crop $-$ annual precipitation | grasses. Bromegrass and intermediate wheatgrass |
| | water tolerance (in) | became the dominant species, not necessarily due |
| | Minimum dimension of one half he needs of food lat with a | to effluent application. Grazing cattle did not dis- |
| • | suggested sizing procedure of: | criminate between areas receiving effluent and |
| - | Travel time should be proportional to BOD concentration | area receiving only water for imgation |
| - | mayer and should be proportional to DOD concentration | |
| • | Distribution lines longer than 30 m created challenges with | • Daily application of waste resulted in tall fescue |
| | uniform flow | being replaced by barnyard grass in early season |
| • | Filter area designed for flow of 4.5 L/m ² VTA/day was a safe | and crabgrass later in the season |
| | VTA was common | monthly basis was preferable to pasturing |
| | | • Duplicate VTA area was needed to allow soil dry- |
| | | ing and harvesting due to daily effluent additions |
| | | High rate "dosing" with a pump was found to he preferable for even distribution and to avoid |
| | | freeze up problems during winter operation |
| • | VTA should be located at least 3 m (10 ft) above ground water or | Quality of vegetation impacts nutrient removal of |
| | seasonal perched water table and 30 m (100 ft) from wells | vegetation. Establishment procedures and har- |
| • | Sedimentation structure must preceed VTA | vesting frequency is important to establishing lush |
| • | For finishing cattle 1 ha of VTA is suggested per 200 head. For | Iorage growth |
| | calves confined for 150 d/yr, 1 ha of VTA is suggested per 1,000 head | |
| • | VTA systems should be sized by matching normal nutrient runoff | |
| | and crop nutrient utilization | |
| • | USDA SCS design specification to pass the peak discharge of a | Preferential flow path from the lip spreader through |
| | 2-yr, 24-n storm at a maximum flow depth of 1.3 cm with a detention time of 15 min was inadequate | the VIA was another identified cause of poor perfor- mance |
| • | A mean hydraulic retention time of 5–8 min within the settling | • Cross drainage across lots should be avoided to |
| | basin was used for peak runoff rates | prevent one area of settling basin collecting most |
| • | Earth bottom settling basin was designed to be cleaned with | solids. Berms or wooden planks at the fence line |
| | iront-end loader. For wet years, a settling basin slope (6 to 1) was selected to allow how scraper to be backed into settling basin | Detween pens were suggested Solids accumulation at the bottom and of the pons |
| | while keeping tractor on dry ground | (due to animal traffic and solids settling) created |
| • | Settling basin drainage to minimize liquid depth was recommended | problems with uneven flow into the settling basin. |
| | to minimize seepage below the basin | Periodic solids removal from under the fence line |
| • | Settling basin outlets were installed to place and maintain all outlets on an equal elevation (reinforced concrete pade set outlet | at the lower end of the feedlot is needed |
| | elevation | considered adequate |
| • | Settling basin drain pipes (separate from normal outlets) were | • Herbicides were used for broadleaf weed control |
| | installed to allow complete basin drainage and solids drying prior to | on the VTA and settling basin berm |
| | solids removal | |

• A design procedure was developed by NRCS in Pennsylvania suggesting that the VTA be designed for the peak discharge resulting from a 2-year, 24-hour storm event at a maximum flow depth of 1.3 centimeters with a minimum flow through time of 15 minutes (Murphy and Bogovich 2001). A design procedure based upon a sheet flow equation was proposed:

$$T = \frac{0.07 (nL)^{0.8}}{(P_2)^{0.5} s^{0.4}}$$

where:

- T = travel time (h)
- n = Manning's roughness coefficient (0.24 for dense grass)
- L = flow length (ft)

 $P_2 = 2$ -yr, 24-h storm

s = land slope (ft/ft)

Scheilinger and Clausen (1992) used this design standard for Vermont applications and observed poor performance results. Additional design criteria have been assembled by other NRCS state offices including the Montana Supplement to chapter 10 of the Agricultural Waste Management Field Handbook (Montana NRCS 2003). All of these practice standards have typically targeted non-CAFO units. For example, the Montana practice standard states that final designs for feedlots larger than 3 acres (about 600 cattle) should not be designed with the simplified method (Montana practice standard).

• Murphy and Harner (2001) suggested sizing a VTA area based upon normal nitrogen runoff balanced against nitrogen removal as harvested hay. Procedures for estimating mass of nitrogen runoff from the feedlot and example design calculations are provided in section 6.

• Black (1984) proposed a design procedure based on a maximum allowable hydraulic load to the filter.

$$R_w = P + \left(\frac{D}{10}\right) + SR$$

where:

- $R_{\rm w}~$ = maximum allowable wastewater hydraulic load in cm/yr
- P = soil permeability in cm/yr
- D = soil water deficit in mm/yr
- SR = seasonal runoff rate in cm/yr
 - After calculating R_w , a required VTA area can be calculated by dividing the total flow expected, which includes wastewater, runoff, and direct precipitation, by R_w .
 - Overcash (1981) proposed a design equation based on influent and effluent concentrations.

$$\mathbf{C}_{\mathrm{X}} = \mathbf{C}_{\mathrm{B}} + \left(\mathbf{C}_{\mathrm{O}} - \mathbf{C}_{\mathrm{B}}\right) \times \mathbf{e}^{\left\{ \left[\frac{1}{(1-D)}\right] \times \ln\left[\frac{1}{(1+K)}\right] \right\}}$$

This procedure requires knowledge of the influent contaminant concentrations, C_O , to the VTA. A desired VTA effluent concentration, C_X , can then be selected. C_B represents the background concentration, D is the ratio of infiltration to runoff, and K is the ratio of VTA length to waste area length. Once C_X , C_B , C_O , and D have been determined, the equation must be solved for K to size the filter strip. This calculation should be made for all contaminants of concern, and filter strip length be selected based on the limiting contaminant.

VTA maintenance

Section 9

Several maintenance issues are critical in VTA function (table 9–12):

- A good stand of dense vegetation is needed. Dickey and Vanderholm (1981) noted that dormant residues are effective for filtering and settling pollutants. Management practices that contribute to strong fall growth and well-established winter vegetative cover are critical. Regular harvesting (including hay removal), prevention of channel flow, and minimizing solids accumulation in the VTA are of value in achieving dense fall vegetation. Soil testing to determine fertilization will be of value.
- Uniform flow conditions are essential to VTA performance. Minimal animal traffic and limiting of vehicle traffic to dry conditions are critical.
- Prevention of nutrient accumulation in VTA is important. Regular harvesting with crop removal to encourage a balance of nutrients of nutrients is necessary. Animal grazing is not an acceptable harvesting option. Regular soil testing for residual soil nitrates and phosphorus is suggested at the upper end of the VTA. Higher nutrient deposition is anticipated in the first few meters of the VTA suggesting a potential for nitrate leaching and increased soil P.

Conclusions

Based upon this literature review, the following conclusions are drawn about the application of vegetative treatment areas to runoff from open lot livestock production systems:

- Substantial research (approximately 40 identified field trials and plot studies) provides a basis for understanding the performance of VTS. A superior research knowledge base exists for performance of VTS as compared to baseline systems for CAFO regulation compliance.
- The baseline systems for CAFO regulation compliance perform well in the High Plains regions of the United States where significant moisture deficits exist (rainfall minus evaporation). However, the performance of these baseline technologies drops substantially for decreasing moisture deficits found in the central and eastern Corn Belt states. These trends have been established through computer modeling processes. In-field performance measurements do not exist for baseline systems established by CAFO regulations.
- The existing research targeting VTS is confined to non-CAFO applications, likely due to past regulatory limits. Unique challenges exist in adapting these results and recommendations to CAFO applications.
- The pollutant reduction resulting from a VTS is based upon two primary mechanisms: sedimentation, typically occurring within the first few meters of a VTS, and infiltration of runoff into the soil profile. Systems relying primarily on sedimentation only are unlikely to perform equal or better than baseline technologies. System design based upon sedimentation and infiltration is necessary to achieve a required performance level for CAFO application.

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